A NOVEL USE OF CALCIUM ALUMINATE CEMENTS FOR RECYCLING WASTE FOUNDRY SAND (WFS)


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ABSTRACT

The recycling of WFS in calcium aluminate cement mortars at high level of replacement (50%) was studied. Consistency and setting time were measured in fresh mortars. Two different curing conditions were used, the second one with the aim of enhancing the formation of stable cubic calcium aluminate hydrates. Compressive strength, pore size distribution and SEM examination were carried out to assess the effectiveness of CAC matrix in the recycling of WFS. The effect of the incorporation of a polymeric admixture – a chitosan derivative – was also explored. Leaching studies on hardened mortars of three target pollutants of the WFS (toxic metals Pb, Cr and Zn) were also done to evaluate their immobilization. By comparison with OPC mortars, the use of CAC showed several advantages, improving the compressive strength and the toxic metal retention.

Keywords: calcium aluminate cement; leaching; compressive strength; SEM; toxic metals; pore size distribution.
1. INTRODUCTION

Foundry sand is a by-product of ferrous and nonferrous metal casting industries, in which sand is used as a molding material. When the sand can no longer be reused, it becomes a hazardous material according to European regulations as quoted by Alonso-Santurde et al. [1], since it can contain heavy metals and different pollutants [2-3]. The usual fate of this sand – several MT per year - is to be landfilled after a previous process of solidification/stabilization. Owing to high landfilling costs and environmental concerns, some researchers have reported other management alternatives for these polluted WFS, such as subbases, soils, ceramics, controlled-low-strength materials and concretes [4-8]. In these last two cases, WFS can be used to partially replace normal sand. Different issues related to this last point have been addressed: the effect of WFS on workability and compressive strengths, the freezing-thawing resistance of the concretes and the replacement levels [9-11]. In general, only low replacement levels (below 20%) guarantee acceptable performance in the final concrete [5].

All the preexisting work shows two common characteristics: on the one hand, ordinary Portland cement (OPC) was used as the binding material; on the other hand, the recycling was focused on the preparation of concrete (with gravels, particles of aggregate showing diameter > 5 mm) [4,5,10,11]. However, the present study aims to draw attention to: (i) the use of calcium aluminate cement (CAC) as an alternative binder to effectively solidify WFS, and (ii) the obtaining of mortars instead of concrete, with many different applications other than the structural concrete (unlike concrete, mortars are made of just fine sands). CAC is a kind of cement that is usually employed in the production of fire-resistant materials and also when concreting in temperatures below zero or when a fast increase in strength is required [12]. High early strength, chemical attack resistance and abrasion endurance, among other factors, have been
reported for CAC-based materials [13], thus improving the encapsulation of certain toxic and radioactive wastes [14]. The clinkerization of a calcium aluminate cement yields anhydrous compounds, such as CA, CA$_2$ and C$_{12}$A$_7$, whose reactivity increases as CaO content rises. The CAC reaction with water usually leads to the formation of metastable hexagonal phases, CAH$_{10}$ and C$_2$AH$_6$ and amorphous aluminium hydroxide, which is followed by the formation of the stable cubic phases C$_3$AH$_6$ and AH$_3$ at a higher temperature and in the presence of humidity. The latter usually involves a porosity increase – as a consequence of a contraction in the volume of solids – and a subsequent decrease in strength [15].

The use of CAC for WFS recycling and studies on the obtaining of mortars by sand replacement have so far not been reported. Therefore, this work focuses on the re-usage of WFS in CAC-based mortars with a high replacement level of the regular sand (50%). The effect of the WFS on the CAC-based mortar properties (workability, setting time, compressive strength) are discussed, together with a complete microstructural characterization. A comparison with an OPC-based mortar is also presented. Additionally, a batch of fresh mixtures was modified upon the addition of a biodegradable polymer, carboxymethylchitosan –CMCH– which had successfully been used to improve the retention of toxic metals [16]. Finally, leaching studies were also conducted in order to ensure proper encapsulation of the toxic metals found in the WFS.
2. MATERIAL AND METHODS

2.1. Materials

An iron-rich calcium aluminate cement (CAC) (provided by Ciments Molins) and an ordinary Portland cement (OPC) (CEM II 32.5 N, supplied by Portland S.A. Olazagutía, Spain) were used as binders for the preparation of the different batches of mortars. As aggregates, a regular sand of known grain size distribution (standardized siliceous aggregate, supplied by Eduardo Torroja Institute) and a waste foundry sand (supplied by Electroaceros, S.A., Pamplona, Spain) were used. The grain size distributions and chemical compositions of the sands were as described in Fig. 1 and Table 1, respectively. The amount of toxic metals was higher in WFS than in the regular sand (Table 2). For all the batches tested a constant water:cement ratio, 0.37:1 was used throughout.

One of the batches of samples was also modified by the incorporation of a polymeric admixture, carboxymethyl chitosan (CMCH), in a 0.013 wt. % with respect to the binder.

Table 3 shows the weight proportion of each one of the batches of samples. A total of 4 batches of samples was prepared: samples A made by just CAC with regular sand that acted as a control for CAC-bearing samples; samples B and C, obtained with CAC and aggregate made of 50% regular sand and 50% WFS. In samples C, the CMCH is also present as admixture. Finally, samples D were made of OPC and sand (mixture of 50% regular sand and 50% WFS). Therefore, except for samples A, in all instances, regular sand was replaced by WFS for 50% of the weight.
2.2. Mortar preparation

Dry raw materials, in the ratios reported in Table 3, were blended for 5 min with a BL-8-Ca (Lleal, S.A.) mixer. Water was then added and mixed for 90 s at low speed in an Ibertest IB32-040E apparatus.

Cylindrical 40 mm height x 37 mm diameter specimens, prepared in triplicate, were cured under two different curing conditions:

- Curing condition 1: 20º C and 95% RH [17] over the whole curing period.
- Curing condition 2: 60ºC and 100% RH [18] for 24 hours. Once this period was completed, temperature and RH were shifted towards 20ºC and 95% RH.

Stable hydrated calcium aluminate phases were expected to develop easier under curing condition 2.

2.3. Experimental methods

For the fresh state properties of the mortars, different standardized tests were used to measure consistency, water retention capacity and setting time [19-21].

Once the samples had been cured for 28 days, compressive strength tests were carried out on a Proeti ETI 26.0052 at a loading rate of 50 N s⁻¹ [22]. Pore size distributions (PSD) were obtained by mercury intrusion porosimetry (MIP) using a Micromeritics AutoPoreIV9500 with a pressure ranging between 0.0015 and 207 MPa. X-ray diffraction (XRD) was used to study the mineralogical composition in a Bruker D8Advance, with a CuKα₁ radiation (step size 0.02° (2θ) step time 1 s; from 2° to 50° (2θ)). A semi quantitative approach based on the diffraction peaks intensities was
applied with the aim of establishing the relative amount of the different crystalline phases [23].

The textural characteristics of the samples were observed, after coating with a palladium-gold film, by means of scanning electron microscopy (SEM), in a Zeiss microscope, coupled to a EDS detector.

In samples B, C and D, leaching tests of heavy metals were done. Similar cylindrical test pieces obtained as described above were placed in sealable 110 mm x110 mm x110 mm methacrylate tanks. The determination of the leaching of inorganic components of mortars was carried out following the Tank Test [24]. This test requires monolithic hardened samples to be placed in 1 L of demineralized neutral pH water (leaching fluid). The leachate was replenished at specified times.

After the indicated period, all the leachate was drained off and filtered through a 0.45 μm-pore size syringe filter (CHMLAB Group). Conductivity and pH measurements were carried out by means of a 4-Star benchtop pH/conductivimeter. A 10 mL aliquot of each sample was stored in sub-boiling HNO₃ in order to quantify the leached metals. Eluates were collected at eight different times (0.25, 1, 2.25, 4, 9, 16, 36 and 64 days).

Quantification of metals was accomplished by AAS following the reported experimental conditions [16].

From the data collected, the measured cumulative leaching for each component was calculated using the procedure described in the norm for leaching of monolithic materials [24], recently used and reported elsewhere [25]:
Leaching mechanism was determined from the slope of the linear regression analysis (log–log plots of cumulative leaching vs. time) through the eight data points, considered in different increments of time, as detailed in the EA NEN 7375:2004 norm [24].

3. RESULTS AND DISCUSSION

3.1. Consistency and setting time of the fresh mortars

Assays carried out in the CAC mortars loaded with WFS (samples B and C) showed (Table 4) that incorporation of WFS caused a decrease in the slump as a consequence of the different grain size distribution of the polluted sand, which shows a unimodal distribution centered around 0.25 mm with a smaller ratio of both the finest (≤ 0.125 mm) and largest (≥ 0.5 mm) grains, as seen in Figure 1. The presence of an outstanding number of fine particles (≤ 0.25 mm) caused a decrease in the fluidity of the fresh mixtures, in line with previously reported data [26]. It must be considered that the fineness modulus of the WFS was 2.35, in the range of Medium Fineness modulus according to the EN 13139 norm [27]. However, the regular sand showed a higher fineness modulus (2.85), just in the limit of the Medium Fineness range. In the end, WFS meets the provisions of the standard on sand for masonry mortars, showing a particle size ≤ 2 mm [28].

On the other hand, the replacement of standardized sand by WFS (samples B and C) induced a marked decrease in the setting time with respect to the regular sand CAC mortar (samples A). This is an especially favorable feature from a practical point of view when using these mortars to encapsulate the wasted sands.
As for the effect caused by the presence of carboxymethyl chitosan (CMCH) (mortars C), the most beneficial feature is the marked decrease in the setting time (42%) that results in a speedier solidification/stabilization of the WFS (Table 4). This reduction in the setting time as a consequence of the action of the CMCH is in good agreement with the behavior also reported for OPC-based mortars [29].

3.2. Mechanical, mineralogical and microstructural characterization of the 28-day cured mortars

Hardened specimens after 28 days of curing were subjected to mechanical tests. Figure 2 shows the compressive strength values attained for the two curing conditions used.

It is clear that, after curing condition 1, compressive strength of both CAC and OPC mortars loaded with 50 wt.% WFS (samples B, C and D) diminished with respect to the CAC made up with only regular sand (samples A). The different arrangement of the particle size distribution (PSD) when WFS was added explains these results. As a matter of fact, WFS showed a larger amount of fine particles (≤ 0.25 mm) and a smaller amount of bigger particles (≥ 0.5 mm) than those of the regular sand. Consequently, on the one hand the PSD packing resulted in a low degree of compactness and, on the other hand, the presence of a large population of small particles hindered a good adherence at the interface between the binding material and the larger aggregate particles. These two factors had a synergistic impact on the mechanical resistance, explaining its reduction. Nevertheless, it is also noteworthy that in samples under condition 1 of curing:

(i) CAC mortars with WFS (samples B and C) showed a much larger mechanical strength than the OPC mortars with WFS (sample D), and
(ii) CAC specimens loaded with WFS (B and C) exhibited compressive strengths that could be classified as the most resistant M-20 mortars, according to the specifications for masonry mortars [30].

These facts prove CAC mortars to be an excellent end destination to conclude the usable life-cycle of foundry sands, presenting improvements with respect to the OPC mortars.

In CAC-based systems, it is well-known that, over the years, the conversion reaction from hexagonal metastable hydrates to the cubic stable ones causes a decrease in the mechanical resistance. Assays in a more drastic set of curing conditions (condition 2, 60ºC/100% RH) were used in order to estimate the long-term mechanical behavior of the CAC mortars. In fact, these curing conditions showed that conversion took place and cubic stable calcium aluminate hydrates were found by XRD, as depicted in Fig. 3. These stable hydrates are denser and, accordingly, render the cementitious matrix more porous and less resistant [31]. The forced conditions lead to a drop in the compressive strength that is expected to represent the maximum compressive strength loss over the years. In spite of this, the strength values measured for WFS-loaded CAC (B and C) fulfilled the strength requirements of the specifications for mortars for masonry (all CAC mortars subjected to curing condition 2 were M-10 mortars) and can be used for many different applications [30]. Simultaneously, the measured strengths remained well above the limits for landfilling [32].

XRD analysis led to the characterization of the chemical forms of the calcium aluminates present in samples B and C. Owing to the complexity of the CAC matrices, a semi-quantitative approach, based on the intensities of the diffraction peaks, was used to assess the presence of the phases and their relative amount (Fig. 3).
Monocalcium aluminate, CA (CaAl₂O₄, main peaks: 30.1°, 19° 2θ), appeared as the main anhydrous compound. This phase is predominantly formed after the clinkerization during the CAC production. Also metastable hydrates (CAH₁₀ peaks: 12.4°, 25.06°, 27.3° 2θ and C₂AH₈ peak: 18° 2θ) and, in some samples, cubic stable compounds (C₃AH₆, hydrogarnet, peaks: 17.26°, 31.8°, 33°, 44.46° 2θ and AH₃, gibbsite, peaks: 18.33°, 20.3°, 36.6° 2θ) were identified (Figure 3). Chitosan-free specimens (sample B) cured under condition 1 showed mainly anhydrous and metastable compounds. After performing curing condition 2, these samples allowed us to observe cubic stable hydrated phases. The presence of the CMCH polymer (sample C) enhanced the formation of the stable compounds, even under curing condition 1.

Pore size distribution assessment carried out by MIP showed, in CAC-mortars, different pore size distribution patterns as a function of the prevailing presence of either metastable or stable calcium aluminate compounds. The main presence of metastable compounds (CAH₁₀ and C₂AH₈) led to a unimodal pore size distribution, with pores of small diameter and main pore diameter at ca. 0.1 μm. This fact can be observed in the pore size distribution of samples A and B, after curing condition 1 (Fig. 4 a and b).

However, when conversion reaction took place (curing condition 2), the prevailing amount of the denser stable cubic calcium aluminate hydrates caused a porosity increase, with a shift of the main pore diameter towards higher diameters (ca. 1 μm), that justified the compressive strength drop (see the pore size distributions of samples A and B after condition 2 of curing, Fig. 4 a and b).

MIP of CMCH-modified mortars (samples C) confirmed the XRD results, thus proving that the addition of the chitosan derivative as admixture enhanced the formation of stable calcium aluminate hydrates. In fact, the pore size distribution of samples C
depicted the pattern of stable compounds, even after condition 1 of curing, with main pore diameter at ca. 1 μm and reduced porosity below 0.1 μm. Fig. 4 c shows the pore size distributions of samples C, after the two curing conditions, yielding similar curves.

The replacement of the regular sand by WFS caused an increase in the large porosity (pores above 1 μm), owing to the aforementioned poor degree of the packing of the aggregate particles when WFS was present.

Microstructural examination of the samples by SEM showed that the sand particles of the WFS were well embedded in the CAC-cementitious matrix. Micrographs of samples B, under condition 1 of curing, depicted some areas showing a compact matrix, with gelified metastable hydrates (Fig. 5a). Sometimes cracks also were identified over the surface of the sample (Fig. 5b and c). At higher magnification, an ill-defined structure of hydrates could be observed, with some rods and plate-like crystals (Fig. 5d). SEM examination confirmed that these samples are of overall reduced porosity, as discussed in the MIP results.

For these samples B, cured under condition 2, a larger ratio of porosity and a pore size increase was observed as a consequence of the conversion into stable hydrates, as seen in the micrographs at low magnification (Fig. 6a). At higher magnification, the gel structure appeared to be less compact than the corresponding sample under condition 1 of curing (Fig. 6b). Some areas of these samples showed tabular gibbsite crystals together with some needle-like stable hydrates and certain hexagonal plate-like metastable hydrates that were also identified by XRD as C₂AH₈ compounds (Fig. 6c). A detail of tabular agglomeration of stable hydrates is depicted in Fig. 6d. In Fig. 6e, spherical agglomerates of AH₃ of ca. 1 μm within AH₃ gel could be neatly distinguished
[33], whereas in Fig. 6f, C₃AH₆ prismatic and tabular clusters of crystals of stable hydrates were seen.

The incorporation of the chitosan derivative (samples C), after condition 1 of curing, gave rise to a less homogeneous matrix when compared to the chitosan-free samples B. MIP results, above discussed, indicated larger porosity and bigger pore size for these samples, as a result of the conversion into stable hydrates that was enhanced by the polymeric admixture. Some pores and a lower degree of homogeneity in the matrix could be observed in Fig. 7a. An ill-defined structure, poorly compacted, with some quasi spherical AH₃ crystals was seen in Fig. 7b. XRD results suggested the formation of stable cubic hydrates in these samples: this fact was confirmed by the presence of cubic and prismatic C₃AH₆ compounds, together with clusters of spherical AH₃ crystals in AH₃ gel matrix, as shown in Figs. 7c-f. In Fig. 7d and Fig. 7e, also hexagonally-shaped pinacoids related to the remaining presence of C₂AH₈ could be identified. Finally, Fig. 7f depicted also very thin sheets of stable hydrates.

Chitosan-bearing samples C after curing condition 2 presented similar microstructural features to samples subjected to curing condition 1, in good agreement with the MIP studies. Micrographs present a porous matrix accompanied by a low degree of compactness (Fig. 8 a-c), with some cracks on the surface of the gel compounds. Besides a poorly crystallized matrix, some cubic C₃AH₆ crystals could also be observed (Fig. 8d). Sometimes the hydrates showed a foil-like morphology, as depicted in Fig. 8e and detailed at higher magnification in Fig. 8f.

When the binder was OPC (samples D), under curing condition 1, the matrix showed reduced porosity (Fig. 9a and b). In some areas, fibrous C-S-H compounds and needles of ettringite were observed (Fig. 9c and d). Small spherical particles, probably fly ashes
belonging to the OPC cement added during the cement production, could be identified (Fig. 9e). The curing under condition 2 for OPC-based mortars yielded an irregular matrix comprising some well-crystallized areas as well as other ill-defined structures (Fig. 9f).

To summarize, from a mechanical and microstructural point of view, the high level of replacement of regular sand by WFS in calcium aluminate cement matrices showed certain advantages in comparison with the same level of replacement in ordinary Portland cement systems. Although the addition of a polymeric chitosan derivative – CMCH- as admixture in low amounts provoked a slight compressive strength reduction after condition 1 of curing, the presence of this admixture guaranteed that stable phases were achieved in short periods of curing at low temperatures without risk of later conversion and subsequent drop in mechanical resistance of the CAC mortar at longer periods of time. Additionally, the CMCH incorporation, at such a small percentage as 0.013%, to CAC mortars loaded with WFS conferred a plus that ensures the quick stabilization of the wasted sand within the mortar owing to the setting time reduction. Under condition 1, these CAC mortars with high replacement level of WFS (samples B and C) showed a compressive strength that, while inferior to the samples A (only regular sand), was above the strength for a M-20 masonry mortar [30]. On the other hand, the incorporation of 50 wt % WFS into OPC (samples D) resulted in a weaker structure with a much more reduced compressive strength (just a M-10 mortar). This kind of CAC-based mortars could be useful for the obtaining of technical refractory mortars, quick bricklaying, refractory brick joining or corrosion resistant mortars.
3.3. Leaching tests

Given that WFS appeared to be polluted by different heavy metals, three of them (Pb, Zn and Cr) were used as targets to monitor the retaining capacity of the mortars assessed. The selection was done on the basis of their relevance as toxic metals and their pernicious influence on the physical properties of mortars. Figure 10 depicts the comparative percentages of metals retained-released from the WFS when incorporated into mortars (samples B, C and D). Pb was seen to be fully retained by all the three batches of mortars, while Cr and Zn underwent a certain release from all three matrices (although in relatively small quantities). Nevertheless, maximum retentions for the metals were always achieved for the CAC mortars that had been modified by the addition of the chitosan derivative (samples C), whereas OPC mortars (samples D) showed the worst retention capacity in all instances. This circumstance supports the addition of CMCH as an effective retaining agent for heavy metals useful to stabilize/immobilize them into cement mortars, as previously described [16].

Conductivity measurements in Fig. 11 carried out in the eluates over time showed parallel patterns for the three assayed specimen types (samples B, C and D), in which the maximum conductivity is always obtained for the leachates of the OPC mortars (samples D). This fact is in good agreement with the highest portion of metals being released from that kind of mortar. Mechanistic approaches on the leaching process were based on the slopes of the log-log plots of derived cumulative leaching vs. time for different time increments, according to the norm [24]. The criteria reported for the interpretation were: when slopes were below 0.35, a surface wash-off (in the case of increment 1-4 from 0.25 to 4 days) or depletions (for all the rest of time increments) prevailed; if gradient was between 0.35 and 0.65 the proposed mechanism was diffusion (for all the studied time intervals), whereas values higher than 0.65 indicated delayed
diffusion (interval from 0.25 to 4 days) or dissolution (for all the rest of time increments). Table 5 displays the slopes calculated and associated standard deviations for the leaching of Al and Fe as tracers of the stability of the CAC matrix as well as of the toxic metals Cr and Zn, found in WFS. When attention was focused on the leaching behavior of Al, it was seen that in the long-term (time increment 5-8: 9-64 days) OPC mortars (samples D) yielded a larger slope for the release of Al than CAC mortars (samples B and C). In this case, while a depletion model was assigned to the Al within the CAC mortars, a diffusion pattern was ascribed to the OPC mortars (samples D). This experimental fact supports the comparatively long-term stability of CAC matrix with respect to the OPC mortars and agreed well with the higher conductivities observed for the OPC mortars. As for Fe, in the global increment (1-36 days), this stability trend could also be observed. In spite of the fact that the CAC matrix presented a larger amount of Fe in its composition, the leaching was diffusion-controlled unlike the OPC mortars (samples D) in which the leaching mechanism was governed by dissolution.

The mechanistic studies for Zn and Cr (Pb not considered since it was completely retained) were hampered by their low concentrations in the leachates owing to the strong retention of these metals within the cementitious matrices. Data in Table 5 suggest that Zn, in general, suffered a diffusion-controlled leaching. However, as for Cr, depending on the curing condition, either a global wash-off or dissolution patterns were followed for CAC mortars, whereas a dissolution mechanism was the dominant pathway in the case of OPC mortars (samples D).
4. CONCLUSIONS

Calcium aluminate cement has been proved to be a suitable matrix to prepare mortars for recycling of waste foundry sand in high level of replacement. As compared to ordinary Portland cement, parameters such as compressive strength and heavy metal retention were markedly improved when CAC was used as binding matrix. In all cases the minimum strength requirements for landfilling (1 MPa) were well complied with. All the tested CAC mortars were classified at least as M-10, taking into account the specifications for masonry mortars (all the tested mortars surpassed the compressive strength value of 10 MPa). Among toxic metals present as pollutants in WFS, Cr and Zn were seen to be better immobilized in the CAC mortars.

Both the compressive strength and toxic metal retention ability of the CAC mortars remained within acceptable ranges after the two test curing conditions followed in this study. The second condition (60°C/100% RH) induced the formation of cubic stable hydrates, as proved by XRD and SEM examination. In these samples, porosity increased with a subsequent strength drop, which never fell below 12 MPa.

Several advantages were drawn from the incorporation of a polymeric chitosan derivative, carboxymethyl chitosan: (i) setting time reduction, with an ensuing enhancement of the WFS’ pollutants retention; (ii) acceleration of stable hydrates formation even under curing condition 1 (20°C/95% RH), while keeping reasonable mechanical resistance; and (iii) better toxic metal solidification/stabilization, especially for Zn and Cr.

This work explores a novel and promising alternative that consists of using CAC as a valuable binding matrix to recycle WFS and other polluted materials. The final purpose of these materials could be to be landfilled or to use them as masonry mortars.
Acknowledgments

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References


Table 1. Chemical composition of major components present in the regular and waste foundry sands.

<table>
<thead>
<tr>
<th>Constituent (%)</th>
<th>SiO$_2$</th>
<th>Al$_2$O$_3$</th>
<th>TiO$_2$</th>
<th>CaO</th>
<th>MgO</th>
<th>Fe$_2$O$_3$</th>
<th>Na$_2$O</th>
<th>K$_2$O</th>
<th>Cr$_2$O$_3$</th>
<th>MnO</th>
</tr>
</thead>
<tbody>
<tr>
<td>Regular sand</td>
<td>93.4±2.1</td>
<td>0.8±0.4</td>
<td>0.73±0.01</td>
<td>0.04±0.03</td>
<td>0.04±0.03</td>
<td>0.06±0.01</td>
<td>0.22±0.01</td>
<td>0.43±0.03</td>
<td>0.001±0.001</td>
<td>&lt;LD</td>
</tr>
<tr>
<td>Waste foundry sand</td>
<td>88.5±1.0</td>
<td>2.7±0.3</td>
<td>0.61±0.02</td>
<td>0.16±0.06</td>
<td>0.19±0.04</td>
<td>1.33±0.02</td>
<td>0.46±0.06</td>
<td>0.77±0.07</td>
<td>0.045±0.002</td>
<td>0.014±0.007</td>
</tr>
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</table>

Table 2. Chemical composition reflecting the minor elements present in the tested sands.

<table>
<thead>
<tr>
<th>Constituent (mg/kg)</th>
<th>Zn</th>
<th>Cu</th>
<th>Pb</th>
<th>Ni</th>
<th>Cd</th>
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<tr>
<td>Regular sand</td>
<td>&lt;LD</td>
<td>4.0±0.9</td>
<td>4.7±0.8</td>
<td>2.4±0.3</td>
<td>&lt;LD</td>
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<tr>
<td>Waste foundry sand</td>
<td>42.9±1.6</td>
<td>18.3±6.5</td>
<td>13.2±2.0</td>
<td>4.3±0.7</td>
<td>0.04±0.01</td>
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Table 3. Weight composition (g) of every batch of mortars.

<table>
<thead>
<tr>
<th>Batches of samples</th>
<th>CAC</th>
<th>OPC</th>
<th>Regular sand</th>
<th>WFS</th>
<th>Mixing water</th>
<th>Admixture (CMCH)</th>
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<tbody>
<tr>
<td>A</td>
<td>390</td>
<td>-</td>
<td>610</td>
<td>-</td>
<td>144.3</td>
<td>-</td>
</tr>
<tr>
<td>B</td>
<td>390</td>
<td>-</td>
<td>305</td>
<td>305</td>
<td>144.3</td>
<td>-</td>
</tr>
<tr>
<td>C</td>
<td>390</td>
<td>-</td>
<td>305</td>
<td>305</td>
<td>144.3</td>
<td>0.05</td>
</tr>
<tr>
<td>D</td>
<td>-</td>
<td>390</td>
<td>305</td>
<td>305</td>
<td>144.3</td>
<td>-</td>
</tr>
</tbody>
</table>

Table 4. Fresh state properties of CAC-based mortars (Samples A-C) and OPC mortars (Sample D).

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Sample A</th>
<th>Sample B</th>
<th>Sample C</th>
<th>Sample D</th>
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<tbody>
<tr>
<td>Slump (mm)</td>
<td>173</td>
<td>114</td>
<td>118</td>
<td>144</td>
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<tr>
<td>Water retention (%)</td>
<td>92.5</td>
<td>98.0</td>
<td>93.5</td>
<td>94.8</td>
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<tr>
<td>Setting time (min)</td>
<td>178</td>
<td>108</td>
<td>63</td>
<td>110</td>
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</table>
Table 5. Calculated slopes and associated standard deviations of the log–log plots of derived cumulative leaching vs. time for different time increments.

<table>
<thead>
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<th>Step</th>
<th>Sample B</th>
<th>Sample C</th>
<th>Sample D</th>
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<tr>
<td></td>
<td>Condition 1</td>
<td>Condition 2</td>
<td>Condition 1</td>
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<tr>
<td><strong>Increment 1-4</strong></td>
<td>1.123±0.039</td>
<td>1.151±0.037</td>
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<td><strong>Increment 2-5</strong></td>
<td>1.111±0.063</td>
<td>1.079±0.083</td>
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<tr>
<td><strong>Increment 3-6</strong></td>
<td>0.936±0.057</td>
<td>0.828±0.068</td>
<td>0.929±0.083</td>
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<tr>
<td><strong>Increment 4-7</strong></td>
<td>0.597±0.141</td>
<td>0.500±0.014</td>
<td>0.586±0.131</td>
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<tr>
<td><strong>Increment 5-8</strong></td>
<td>0.318±0.087</td>
<td>0.208±0.069</td>
<td>0.366±0.052</td>
</tr>
<tr>
<td><strong>Global increment</strong></td>
<td>0.859±0.105</td>
<td>0.786±0.115</td>
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<th>Step</th>
<th>Sample B</th>
<th>Sample C</th>
<th>Sample D</th>
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<td>Condition 1</td>
<td>Condition 2</td>
<td>Condition 1</td>
</tr>
<tr>
<td><strong>Increment 1-4</strong></td>
<td>0.482±0.059</td>
<td>0.605±0.115</td>
<td>0.793±0.111</td>
</tr>
<tr>
<td><strong>Increment 2-5</strong></td>
<td>0.501±0.074</td>
<td>0.654±0.139</td>
<td>0.794±0.160</td>
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<tr>
<td><strong>Increment 3-6</strong></td>
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<td>0.407±0.079</td>
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<td><strong>Increment 4-7</strong></td>
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<td>0.194±0.044</td>
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<td><strong>Global increment</strong></td>
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<td>0.465±0.086</td>
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<td>Step&lt;sup&gt;a&lt;/sup&gt;</td>
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<td>Sample C</td>
<td>Sample D</td>
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<tr>
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<td>Condition 1</td>
<td>Condition 2</td>
<td>Condition 1</td>
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<tr>
<td>Increment 1-4</td>
<td>2.690±0.850</td>
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<td>1.590±0.335</td>
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<td>0.350±0.196</td>
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<tr>
<th>Step&lt;sup&gt;a&lt;/sup&gt;</th>
<th>Sample B</th>
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<td>Condition 2</td>
<td>Condition 1</td>
</tr>
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<td>1.109±0.103</td>
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<td>Increment 4-7</td>
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<td>0.726±0.094</td>
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<sup>a</sup>Increment 1-4: from 0.25 to 4 days; Increment 2-5: from 1 to 9 days; Increment 3-6: from 2.25 to 16 days; Increment 4-7: from 4 to 36 days; Increment 5-8: from 9 to 64 days; Global increment: from 1 to 36 days.

<sup>b</sup>no changes of concentration were observed, so the quantified slope was null.
Figure 1. Particle size distribution of regular and waste foundry sands.

Figure 2. Compressive strengths for 28-day hardened specimens cured under condition 1 (20°C, 95%RH) and condition 2 (60 °C, 100%RH).
Figure 3. Relative abundance (intensity given in vertical axis in arbitrary units) of cementitious crystalline phases in samples B and C cured in either condition 1 or 2, as identified by XRD.
Figure 4. Pore size distribution of: a) samples B; b) samples C and c) samples D subjected to curing conditions 1 and 2.
**Figure 5.** SEM micrographs of samples B after curing condition 1: a) gel-like structure of metastable hydrates, b) and c) microstructure showing cracks over the surface (indicated by arrows); d) ill-defined structure with rods and plate-like crystals.
Figure 6. SEM analysis of samples B after curing condition 2: a) porous microstructure (some pores indicated by arrows), b) gel-like structure; c) tabular (arrows) and needle-like (encircled area) gibbsite crystals together with hexagonal plate-like metastable hydrates; d) tabular crystals agglomeration; e) spherical agglomerates of AH3 (arrows); f) prismatic (arrows) and tabular clusters of stable hydrates.
Figure 7. SEM examination of samples C (containing CMCH) after curing condition 1: a) matrix with a low degree of homogeneity; b) ill-defined structure with quasi spherical AH$_3$ crystals; c), d) and e) cubic and prismatic C$_3$AH$_6$ crystals (solid arrows) and spherical AH$_3$ crystals. Hexagonally-shaped pinacoids of C$_2$AH$_8$ (dotted arrows); f) thin sheets of stable hydrates.
Figure 8. SEM micrographs of the microstructure of samples C (chitosan-bearing CAC mortars) cured under condition 2: a) porous matrix; b) and c) gel-like structure with cracks (arrows); d) poorly crystallized matrix with some cubic crystals (arrows); e) and f) foil-like morphology.
Figure 9. SEM images of OPC specimens (samples D) after curing at condition 1: a) and b) compact matrix; c) and d) fibrous C-S-H compounds (solid arrows) and needle-like ettringite crystals (dotted arrows); e) spherical particles attributed to fly ashes. In sample after condition 2 of curing (f) the matrix showed an irregular microstructure.
Figure 10. Percentages of metals retained and released from the 28-day hardened monolithic samples containing 50 wt.% WFS (CAC: samples B and C; OPC: samples D).
Figure 11. Conductivity of the eluates from samples B, C and D cured under both conditions (1 and 2).